Parameter Estimates for Fishes of the Upper Paraná River Floodplain and Itaipu **Reservoir (Brazil)**

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Abstract

Estimates of the growth (K), natural mortality (M), consumption/biomass (Q/B) rate and trophic level (TL) for 35 species in the upper Paraná river floodplain and the Itaipu reservoir (interconnected ecosystems) are presented. A compilation of these biological statistics is made for comparison purposes and some general trends are briefly discussed.

Introduction

Consumption, natural mortality, trophic level and estimates of growth are very important parameters for the construction of Ecopath (trophic) models (Christensen and Pauly 1993). Consumption (Q) is the intake of food by a species over a certain time period. The ratio Q/B (Consumption/ Biomass) has been defined as the number of times a population consumes its own weight in a year (García and Duarte 2002; Pauly 1998a).

According to Allen (1971), usually Natural Mortality (M), in equilibrium conditions, can be considered the production that is the "total quantity of tissue elaborated by a fish population during a stated period of time" (Ivlev 1945). Trophic Level (TL) defines the position of the fish on food web and can be estimated by stomach content analysis. Growth estimates can be described by the von Bertalanffy growth function (VBGF), which is expressed by the parameters K (growth constant) and L_m (asymptotic length).

In order to infer the population response to impacts from two different ecosystems during two periods, the parameters K, L, M, QB and TL were estimated and compared for the 35 more abundant species from the upper Paraná river floodplain and the Itaipu reservoir.



floodplain and Itaipu reservoir.

Material and Methods Study areas

The Itaipu reservoir is located on the border of Brazil and Paraguay. It has an area of I 350 km² and a mean depth of 22 m (mean water residence time = 40 d) (Figure 1). The reservoir was closed in 1982. The Upper Paraná River includes approximately the first one-third of the Paraná river basin with a drainage area of 2.8 x 106 km² and 138 reservoirs with dams greater than 10 m in height. The study area is situated immediately above the Itaipu reservoir and represents as much as one-third of the original floodplain (230 km). It is the last stretch of the Paraná river in Brazilian territory that has not been impounded (Agostinho and Zalewski 1995; Gomes and Miranda 2001; Petrere et al. 2002).

Data sampling

From March 1992 to February 1995, sampling of fish was conducted every three months in the various habitats (river, channels and lakes) in the upper Paraná river floodplain. Gill nets (2-16 cm mesh) and trammel nets (6-8 cm mesh) were used for collecting the samples. Fishing gears were deployed for 24 hours during each sampling period, with inspections in the early morning, dusk and late evening.

At the Itaipu reservoir, species were collected monthly during January 1983 to December 1987 and January 1988 to December 1992. Gill nets (3-16 cm mesh) were deployed for 24 hours during each sampling period (with inspections in the early morning, dusk and late evening) at various sites (lotic and lentic environments; transitional zone).

The fishes were preserved with tricaine or oil of cloves before being sorted by species, counted and measured.

Data analysis

Length-frequency data for species from both environments (reservoir and floodplain) were analyzed. Growth parameters were calculated using VBGF equations:

$$L_{r} = L_{\infty} * (I - e^{-K(t - to)})$$
 (1)

where $L_{\rm t}$ is fish length (cm) at age t, L_{∞} is the Asymptotic Length (cm), K is the curvature of the VBGF or Growth Rate (year⁻¹), and t_0 ("t - zero") is the age intercept where fish age is assumed to be zero.

Parameter K (growth rate) of the VBGF was calculated using length-frequency data and the ELEFAN I routine of FISAT (Sparre et al. 1989; FAO-ICLARM 1996) (and with option to scan for K-values) because in tropical areas it is difficult, if not impossible, to count rings on hard parts (Lizama and Vazzoler 1993). In all cases, Asymptotic Length (L_{∞}) was calculated using maximum length (L_{\max}), i.e., largest individual of the population sample ($L_{\infty} = L_{\max} * 1.05$).

Natural Mortality (M) was calculated by empiric regression (Pauly 1980):

$$M = K^{0.65} * L_{\infty}^{-0.279} *_{T} 0.463$$
 (2)

where M is Natural Mortality (year- 1) and T is a mean temperature (23.4°C).

Palomares and Pauly (1998) proposed using empiric regression to estimate the Consumption/Biomass rate (Q/B) as follows:

$$\log QB = 7.964 - 0.204 \log W_{\infty} - 1.965 \text{ T}' + 0.083 \text{ Ar} + 0.532 \text{ H} + 0.398 \text{ D}$$
 (3)

where QB is Consumption/Biomass rate (year-1); W_{∞} is Asymptotic Weight (g); T is mean annual temperature of water body, expressed as T'=1000/Kelvin (Kelvin = 23.4°C + 273.15), where 23.4 is the mean annual temperature of reservoir and floodplain; and Ar is the Aspect Ratio of the caudal fin (mm²). H and D are related to feeding behavior, i.e., H=1 for herbivores, D=1 for detritivores.

The Ar value was estimated as the average measurement for three or four individuals per species. Trophic Level (TL) for each species was based on an approach used in Ecopath (Christensen and Pauly 1993), i.e., TL is equal to 1 plus the weighted mean of the prey's trophic level. Information

about diet composition is provided in Hahn et al. (1997) and Agostinho et al. (1997).

Paired t-tests were performed to investigate K value differences between time periods within each ecosystem and among ecosystems, separating reproductive behavior (migratory and sedentary) (Zar 1996). Correlation analysis (Pearson r) was performed on all variable combinations. To control Type II error, p values were adjusted using the sequential Bonferroni test correction of Peres-Neto (1999). Overall, correlation coefficients between parameters are very weak and not significant (p<0.001), hence are not discussed in the results.

Results

Table I presents information and related estimated parameters for 35 species for different time periods and ecosystems, totaling 87 fish populations.

The auximetric plot (from the Greek auxesis - growth and metron - measure) shows that the smaller species have high values of K, while it is the opposite with the large ones. This pattern is not associated with the survey period (Figure 2). Paired t-test (n = 8, p < 0.004) demonstrated that K declined between time periods only for sedentary species such as S. borelii, S. spilopleura and T. paraguayensis. In the Itaipu reservoir, K values for sedentary species were more constant and showed no change for nine years after the closure of the dam (n = 10, p<0.89). Other comparisons, such as between ecosystems, time periods in the Itaipu reservoir, time periods of the migratory species, etc., showed no significant differences.

Discussion

The findings of this study are consistent with other studies that have proposed that smaller species have higher values of K (Pauly 1998a), for instance, the same asymptotic size, may be associated with various values of K (Figure 2), showing different strategies of the occupation of

the "available growth space" by species. These patterns are similar to those observed for fish species in other latitudes and ecosystems (Pauly 1998b).

K values decreased among the sedentary species in the floodplain, probably because these populations are affected by an irregular flood regime (Agostinho et al. 1999, 2001) due to many upstream dams (26 reservoirs covering about 100 km²), while reservoir populations live in an environment with more stability.

The relationship assumed by some authors that Trophic Level increases with fish size (Pauly et al. 1998) is not clear for the Paraná species. The correlation (r = 0.36) was disproportionately influenced by the longest species (*P. corruscans* and *P. luetkeni*).

The compilation also indicated that Q/B varies inversely with L_{∞} and Trophic Level (see Table 1). García and Duarte (2002) showed similar trends for Caribbean fishes. Natural Mortality has a high correlation with L_{∞} and K, but since M is calculated by equation 2, these relationships have no biological significance.

Estimates presented here should be useful for construction of Ecopath models, since in tropical regions like Brazil, ecological modeling is at a development stage (Angelini and Petrere 1996; Wolff et al. 2000; Angelini and Agostinho 2005).

Acknowledgements

This study is part of R.Angelini's PhD. dissertation at the NUPELIA (Research Nucleus in Limnology, Ichthyology and Aquaculture) from University of Maringá, Brazil, under the advice of Miguel Petrere Jr, who made valuable comments. K.O. Winemiller and anonymous provided helpful reviews. Carla S. Pavanelli helped with fish taxonomy.

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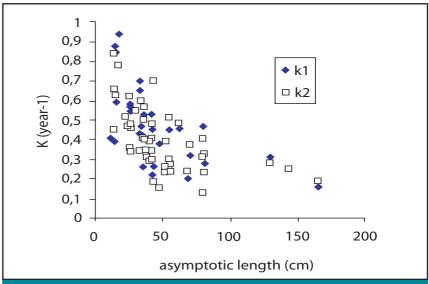


Figure 2. Plot of K and asymptotic length (L^{∞}) (n=87). K1: 1983-1988; K2: 1992-1995 for the Itaipu reservoir and Upper Paraná River floodplain (Brazil).

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Table 1. Estimated parameters for species at	the Itaipu reserv	oir and upper	Paraná river flo	odplain ((Brazil) ir	ı two pe	riods.						
Cuasina	Common	F	Reproductive			I/1	K2	841	Ma	OB	Δ.,	-1	2
Acestrorhyncus lacustris (Lütken, 1875)	name Peixe	Ecosystem Floodplain	behavior Sedentary	TL 3.6	L _∞	K1	K2 0.40	M1	M2 0.9	QB 9.74	Ar 2.73	n1	n2 2371
Acestromyricus lacustris (Lutkeii, 1673)	Cachorro	Reservoir	sedentary	3.6	38	0.40	0.40	0.9	0.9	9.74	2.73	1034	478
Ageneiosus ucayalensis (Casteunal, 1855)	Mandubé	Reservoir	Sedentary	3.7	35	0.26	0.41	0.7	0.9	10.05	2.49	2162	512
Astyanax altiparanae (Garutti & Britski, 2000)	Tambiú	Reservoir Floodplain	Sedentary	2.7 2.7	15 14	0.39	0.66 0.45	1.1	1.6 1.3	12.17 12.17	1.24 1.24	1523	1205 919
Auchenipterus nuchalis (Spix & Agassiz, 1829)	Mandi Peruano	Reservoir Floodplain	Sedentary	3.1 3.1	33 33	0.65 0.7	0.35 0.34	1.3 1.3	0.8 0.8	8.96 8.96	1.85 1.85	5985 791	7306 572
Brycon orbignyanus (Valenciennes, 1850)	Piracanjuba	Floodplain	Migratory	2.8	40		0.29		0.7	6.99	1.6		170
Cyphocharax modestus (Fernández-Yapéz, 1948)	Saguiru	Floodplain	Sedentary	2.1	18	0.94	0.78	1.9	1.7	23.31	0.81	321	304
Hemisorubin platyrhynchos (Valenciennes, 1840)	Braço de moça	Floodplain	Migratory	3.8	52		0.23		0.6	5.81	1.04		211
Hoplias malabaricus (Bloch, 1794)	Traíra	Reservoir Floodplain	Sedentary	3.8 3.8	55 44	0.45	0.51 0.7	0.9	0.9 1.2	6.05 6.05	1.3 1.3	7527	18764 2303
Hoplosternum littorale (Hancock, 1828)	Tamboatá	Floodplain	Sedentary	2.6	23		0.52		1.2	9.10	1.68		1795
Hypophthalmus edentatus (Spix & Agassiz, 1829)	Mapará	Reservoir Floodplain	Sedentary	2.7 2.7	56 56	0.28	0.28 0.24	0.6	0.6 0.6	26.8 26.8	2.04 2.04	7522	18758 171
Hypostomus spp	Cascudo	Reservoir Floodplain	Sedentary	2.0 2.0	41 34	0.53 0.47	0.41 0.6	1 1	0.9 1.2	15.3 15.3	0.94 0.94	465 545	899 553
Iheringichthys labrosus (Lütken, 1874)	Mandi	Reservoir Floodplain	Sedentary	2.7 2.7	33 30	0.43	0.56 0.55	1	1.2 1.2	23.2 23.2	1.63 1.63	1506	3389 1536
Leporinus friderici (Bloch, 1794)	Cabeça Gorda	Reservoir Floodplain	Migratory	2.6 2.6	42 42	0.22	0.48 0.34	0.6	1 0.8	6.38 6.38	1.37 1.37	712	524 864
Leporinus obtusidens (Valenciennes, 1836)	piapara	Reservoir Floodplain	Migratory	2.6 2.6	44 42	0.26	0.18 0.3	0.6	0.5 0.7	6.24 6.24	1.37 1.37	165	207 1297
Loricaria sp	Acari	Reservoir	Sedentary	2.3	36	0.56	0.5	1.1	1	20.9	1.16	115	111
Loricariichthys platymetopon (Isbrücker & Nijssen, 1979)	Acari	Floodplain	Sedentary	2.3	36	0.53	0.57	1.1	1.1	20.9	1.16	7056	5894
Megalancistrus parananus (Peters, 1881)	Cascudo abacaxi	Reservoir	Sedentary	2.0	52		0.26		0.6	10.0			1590
Moenkhausia intermedia (Eigenmann, 1908)	Lambari	Reservoir	Sedentary	3.0	11	0.41		1.3		13.2	0.726	431	
Pimelodus maculatus (Lacepède, 1803)	Mandi	Reservoir Floodplain	Migratory	3.1 3.1	43 40	0.45 0.29	0.41 0.39	0.9 0.7	0.9 0.9	17.8 17.8	1.64 1.64	1223 1057	2710 1213
Pinirampus pirinampu (Spix & Agassiz, 1829)	Pinarambu	Reservoir	Migratory	3.4	68	0.2	0.24	0.5	0.5	6.4	2.48	461	834
Plagioscion squamosissimus (Heckel, 1840)	Pescada branca	Reservoir Floodplain	Sedentary	3.7 3.7	62 48	0.46 0.38	0.48 0.15	0.9 0.8	0.9 0.4	6.0 6.0	1.43 1.43	6623 699	16089 339
Prochilodus lineatus (Valenciennes, 1836)	Curimba	Reservoir Floodplain	Migratory	2.2 2.2	81 81	0.28	0.33 0.23	0.6	0.6 0.5	12.8 16.3	2.16 2.16	2093	2326 6250

Species	Common name	Ecosystem	Reproductive behavior	11	_s°	K1	K2	M1	M2	QB	Ar	n1	
Pseudoplatystoma corruscans (Spix & Agassiz, 1829)	Pintado	Reservoir Floodplain	Migratory	4.0	165 143	0.16	0.19	0.3	0.4	3.7	1.86	250	
Pterodoras granulosus (Valenciennes)	Abotoado	Reservoir Floodplain	Migratory	2.3	70 55	0.32	0.37 0.3	0.7	0.7	5.4	1.35	1603	
Rhaphiodon vulpinus (Spix & Agassiz, 1829)	Peixe- cachorro	Reservoir Floodplain	Sedentary	3.8	80	0.47	0.31	0.8	0.6	5.84	1.76 1.76	1098	,
Rhinelepis aspera (Spix & Agassiz, 1829)	Cascudo preto	Reservoir	Migratory	2.1	53		0.39		0.8				,
Salminus brasiliensis (Cuvier, 1816)	Dourado	Floodplain	Migratory	3.6	80		0.13		0.4	4.4	1.43		
Schizodon altoparanae (Garavello & Britski, 1990)	Piava	Floodplain	Sedentary	2.0	27		0.48		1.1	33.8	2.06		
Schizodon borellii (Boulenger, 1900)	Piava	Reservoir Floodplain	Sedentary	2.0	38 36	0.5 0.53	0.35 0.4	1.1	0.8 0.9	19.8 19.8	2.03 2.03	543 832	. (4
Serrasalmus marginatus (Valenciennes, 1836)	Piranha	Reservoir Floodplain	Sedentary	3.8	26 26	0.58	0.62 0.36	1.3	1.3 0.9	11.5 11.5	3.33	2244 1503	` "
Serrasalmus spilopleura (Kner, 1858)	Piranha	Floodplain	Sedentary	3.8	27	0.55	0.34	1.2	6:0	9.12	2.35	3547	
Steindachnerina insculpta (Fernández & Yépez, 1948)		Floodplain	Sedentary	3.8	14	0.88	0.84	1.9	2	23.7		519	,
Trachelyopterus galeatus (Linnaeus, 1766)	Cangati	Reservoir Floodplain	Sedentary Sedentary	2.6	27 24	0.54	0.46	1.2	1.1	8.92	1.44 44.1	2634	(4
<i>Trachydoras paraguayensis</i> (Eigenmann & Ward, 1907)	Rique-rique	Reservoir Floodplain	Sedentary	2.5	16 15.5	0.59	0.63	1.5	1.5	8.5		2019 638	,
Zungaro jahu (Ihering, 1898)	Jaú	Reservoir	Migratory	4.0	130	0.31	0.28	0.5	0.5	2.5		352	

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806 1618 1962 187 560 462 2869 1363 6723 TL-Trophic Level; L.z. - Asymptotic Length (cm); K. Growth Rate (year¹); M-Natural Mortality (year¹); QB- Consumption/Biomass (t km²year¹); Ar-Aspect Ratio of the caudal fin; n - sample size. Number 1: 1983-1987 for reservoir and 1986-1988 for floodplain; Number 2: 1988-1992 for reservoir and 1992-1995 for floodplain.

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